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**Socio-Ecological Accounting: DPSWR, a Modified DPSIR Framework, and its  
Application to Marine Ecosystems**

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29 **Abstract**

30

31 In contrast to institutionally focussed environmental accounting, socio-ecological accounting  
32 frameworks organise information concerned with human-environment interactions at scales  
33 relevant to ecosystem change and thus encapsulate information more relevant to ecosystem-  
34 based management. The DPSIR (Driver-Pressure-State Impact-Response) framework has  
35 been used to identify relevant information in a number of ecosystem contexts but suffers  
36 limitations in terms of its definitional clarity and conceptual foundations, which undermine  
37 comparability between studies. These limitations are addressed in the DPSWR (Driver-  
38 Pressure-State-Welfare-Response) framework, which defines information categories based on  
39 a synthesis of concepts in DPSIR and its predecessors so as to more clearly identify the object  
40 of measurement in each category and isolate information relating to social systems.  
41 Consequently, its categories dealing with social systems are better suited to assessing  
42 anthropocentric trade-offs in environmental decision-making, such as through cost-benefit  
43 analysis. A conceptual input-output analysis is used to highlight measurement issues  
44 connected with the inter-relations between information categories, particularly as regards  
45 scale, and the application of the framework is illustrated by reference to issues affecting  
46 marine ecosystems included in a Europe-wide study for the European Commission. However,  
47 DPSWR's definitions are designed to be sufficiently general as to support application in other  
48 ecosystem contexts.

49

50 **Keywords:** socio-ecological accounting; environmental accounting; ecosystem approach;  
51 marine ecosystems

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53  
54 **1. Introduction**

55  
56 Environmental accounting involves organising and presenting information on interactions  
57 between the economy and the environment in a standardised way to support policy  
58 development and decision making (UNSD, 2012) and, in the case of corporate entities, as a  
59 basis for social accountability (Gray et al., 1995). This practice is generally associated with  
60 institutional accounting frameworks, where the scope of accounting is determined by the  
61 attribution of environmental resources to socially defined entities. However, indicator-based  
62 frameworks which integrate information on associated changes in social (human) and  
63 ecological systems with a scope relevant to ecological change constitute the basis for socio-  
64 ecological accounting frameworks (Table 1). Thus, they are more appropriate to fulfilling the  
65 information needs of integrated management (Grumbine, 1994; McFadden and Barnes, 2009)  
66 as embodied in the ecosystem approach (see, for example, CBD, 2000; Environment Canada,  
67 2012; EU, 2008).

68

69 **[Table 1 about here]**

70

71 In the simplest conceivable cases, analysis can be confined to a single ecosystem change with  
72 a unique anthropogenic cause at a given location. More generally, socio-ecological  
73 accounting requires information categorised so as to capture multiple causes of ecosystem  
74 change, the nature of that change and the range of effects on social systems in a manner that  
75 supports the analysis of the complex interactions among those categories. The DPSIR  
76 (Driver-Pressure-State-Impact-Response) accounting framework comprises information  
77 categories based on a chain of causal relations that together encompass these phenomena and  
78 has been widely adopted in a variety of ecosystem contexts, for example: soil erosion (Gobin

79 *et al.*, 2004); biodiversity loss (Maxim *et al.*, 2009); and, marine and coastal systems (Bowen  
80 and Riley, 2003; Cave *et al.*, 2003). However, as explained below, the definitions of this  
81 framework's information categories are subject to interpretation and they are not well aligned  
82 conceptually with the economic analysis of interactions between social and ecological  
83 systems.

84  
85 Some flexibility in the interpretation of information categories allows users to adapt the  
86 framework to their particular needs but this flexibility needs to be limited to ensure  
87 comparability between studies, making the accumulation of knowledge about specific social-  
88 ecological interactions more efficient. This issue was highlighted in a recent European  
89 Commission-funded project, ELME (European Lifestyles and Marine Ecosystems), which  
90 motivated the study reported here. The project involved multidisciplinary teams analyzing a  
91 range of ecosystem changes in European seas making it necessary to adopt a standard set of  
92 information category definitions that could be uniformly applied in each case, regardless of  
93 the type of ecosystem change or sea under investigation, and supporting common indicator  
94 measures so that the results could be aggregated. For this purpose, a modified DPSIR  
95 framework (DPSWR, Driver-Pressure-State-Welfare-Response) was designed to improve  
96 definitional clarity and the alignment of social system categories with the needs of economic  
97 analysis. While the DPSWR framework was thus inspired by the needs of the ELME project,  
98 and its application is illustrated here in the context of marine ecosystem degradation, the  
99 definition of terms is sufficiently general for it to be utilised in other contexts.

100  
101 This paper describes the derivation of the DPSWR framework in two stages. First, it  
102 critically reviews the DPSIR category definitions to identify definitional and conceptual  
103 limitations, supported by comparison with predecessor frameworks which have a common

104 conceptual heritage - DPSIR may be seen as the latest generation in an evolutionary process.  
105 In doing so, the paper provides a historical perspective on the origins of current practice,  
106 highlighting the risks of conceptual confusion that have arisen. Second, the results of the  
107 review are synthesised through modifications to DPSIR that result in the DPSWR framework.  
108 A spatial input-output representation of the causal relationships among information categories  
109 is employed to discuss measurement issues in the application of the DPSWR definitions. In  
110 these sections, eutrophication, primarily from agricultural run-off, is used to illustrate certain  
111 points; the subsequent section uses results from a survey of natural scientists in the ELME  
112 project to illustrate other applications. Finally, concluding remarks deal with the advantages  
113 of the DPSWR framework, how its scope can be reconciled with that of institutional  
114 accounting frameworks and aspects of the framework requiring further development.

115

## 116 **2. Review of DPSIR and Predecessor Frameworks**

117

118 The review is structured around the DPSIR information categories as originally defined by the  
119 European Environment Agency (EEA, 1999) or as subsequently modified in its glossary  
120 (EEA, 2012), as summarised in Table 2.<sup>1</sup> This table aligns, as far as possible, these categories  
121 with those employed in predecessor frameworks extending back to the Stress-Response  
122 Environmental Statistical System (S-RESS) proposed by Friend (1979); no earlier  
123 frameworks could be identified in the literature.

124

125

**[Table 2 about here]**

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<sup>1</sup> The European Environment Agency is considered an authoritative source as it or its parent body, the European Commission, have been cited as the source for the framework (e.g. Bowen and Riley, 2003) and the framework has been widely applied by the EEA itself (e.g. EEA, 2007) and in projects for the Commission (e.g. Eurostat, 1999).

## 126 2.1 Overview

127

128 The summary in Table 2 demonstrates the shared heritage of core concepts in terms of the  
129 characteristics of social and ecological systems of relevance, but highlights omissions in  
130 certain categories, most notably the restriction to environmental effects in the FDES  
131 (Framework for the Development of Environment Statistics) and PSR (Pressure-State-  
132 Response) frameworks, and a sub-division of its pressure category in the PSR/E (Pressure-  
133 State-Response/Effect) framework. Nevertheless, a wide variation in terminology and subtle  
134 variations in the precise definition of information categories are apparent. For example, the  
135 S-RESS “stressor” and the PSR “indirect pressure” categories both refer to human activities,  
136 but they most closely align with the Driver concept reflected in DPSIR which refers to large-  
137 scale social “developments”. Thus, there is a common reference to “what people do” but a  
138 distinction between specific activities in S-RESS and PSR, and a more summary  
139 representation based on trends in DPSIR.

140

141 Certain of the frameworks are not characterised solely by the definitions of the constituent  
142 information categories but also by a structuring property: an orientation towards a motivating  
143 object of analysis. S-RESS is additionally structured around “stressor activities”, e.g.  
144 generation of waste residuals, harvesting activity, while affected systems provide the  
145 orientation for FDES, which uses a media, or environmental component, approach (e.g.  
146 atmosphere, water) based on an international survey (UN, 1977). Similarly, PSR is orientated  
147 around “issues” associated with particular environmental media, e.g. climate change, water  
148 resources. However, such structuring properties are more in the nature of guidance for  
149 application of the framework rather than integral to the definition of informational categories.  
150 The following sub-sections consider each of the DPSIR categories in turn.

151

## 152 2.2 Driver Category

153

154 Common to the frameworks summarised in Table 2 is the concept that environmental status or  
155 change is ultimately the result of, or driven by, humans.<sup>2</sup> However, there is little consensus  
156 around how this force should be defined and measured.

157

158 The EEA definition represents Drivers as “developments” in fundamental, broad-scale aspects  
159 of social systems. While this scope is appropriate to some forms of analysis, it does not allow  
160 for the concept of some constant aspect of social systems exerting an influence on ecosystems  
161 and, moreover, involves information that may be too highly aggregated to elucidate the  
162 interactions between social and ecological systems, e.g. in terms of the extent to which they  
163 are coupled. To illustrate in the context of eutrophication, in the EAA scheme one of the  
164 relevant “developments”, and so a Driver, might be an increasing demand for agricultural  
165 output with an associated Pressure being an increased release of eutrophying agents (nitrogen,  
166 phosphorus, potassium), i.e. assuming increased use of man-made fertilisers to meet the  
167 increased demand. This approach would then fail to capture information on the effect of  
168 fertiliser use where there is a constant rate of such use, or indeed varying rates of use where  
169 there is no change in demand for agricultural output, and ignores the potential for decoupling  
170 in the relationship between fundamental economic forces and pressures. For example,  
171 changes in price or technology, or policy measures, may result in decoupling between  
172 fertiliser inputs and agricultural output.

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<sup>2</sup> While the EEA glossary refers to “driving force(s)” as opposed to “driver(s)” as used elsewhere (e.g. Cave *et al.*, 2003), there appears to be no conceptual distinction since the EEA uses the terms interchangeably (EEA, 1999; EEA, 2000).

173  
174 Applications and reviews of the framework in practice, see for example, Cave *et al.* (2003),  
175 Turner *et al.* (2000) and Wieringa (1999), broadly follow the EEA definition and the above  
176 interpretation, although Cave *et al.* (2003) and Turner *et al.* (2000) expand on the nature of  
177 drivers by specifying types of change or processes (e.g. urbanisation, agricultural  
178 intensification). However, the predecessors of DPSIR (Table 2) reinforce the concern that the  
179 framework as defined by the EEA focuses the Driver category at too broad a scale. While  
180 different terms are employed, a common theme in the definition of “stressor” (S-RESS),  
181 “action” (FDES) and “indirect pressures” (PSR and PSR/E) is that they refer to “activities” –  
182 the use of fertilisers in the eutrophication example. Thus, they envisage specific actions as  
183 giving rise to environmental effects. Furthermore, the PSR/E framework draws a distinction  
184 between “underlying” and “indirect” pressures which reconciles the scope of the DPSIR  
185 Driver category to those of other frameworks in that “underlying” pressures correspond to  
186 broad-scale changes in social system as envisaged in the DPSIR Driver category while  
187 “indirect” pressures correspond to specific activities at the source of environmental change as  
188 envisaged in these other frameworks: respectively, the demand for agricultural output and the  
189 use of man-made fertilisers in the eutrophication case.

190

### 191 2.3 Pressure Category

192

193 Consistent with its definition of Driver, the definition of Pressure in the EEA scheme refers to  
194 “developments”, again suggesting the need for some change in rate to constitute a pressure  
195 although, for example, a constant rate of emission can equally well be seen as a pressure on  
196 the assimilative capacity of ecosystems.

197

198 This category is problematical in terms of detecting a common theme in the other frameworks  
199 that might act as a guiding precedent. S-RESS uses the concept of “stress” but defines this in  
200 terms of pressures and FDES does not explicitly include a category corresponding to DPSIR’s  
201 Pressure. However, both PSR and PSR/E regard a pressure as being a manifestation of  
202 human activities that directly acts on the environment (i.e. as “proximate” or “direct”  
203 pressures respectively), the simplest example being emissions. This notion is implicit in the  
204 PSIR and DPSIR pressure categories but by their reference to resource and land use also  
205 encompasses human activities that are elsewhere treated as the equivalent of drivers, creating  
206 a potential conflict between the categories.

207

#### 208 2.4 State Category

209

210 Distinct from the definitions of other categories, EEA’s definition of State refers to  
211 measurement (indicators) rather than simply what is the object of measurement but it is  
212 unclear whether this has any significance. Furthermore, it is unclear whether its reference to  
213 the “condition of *different* environmental compartments and systems” (emphasis added)  
214 envisages that a measure of State must summarise various aspects of the environment related  
215 to a given issue. Together, these observations indicate the need to define State in terms of the  
216 object of measurement (rather than a particular measure) while allowing this object to be that  
217 most relevant to the analysis in hand (rather than to “different ... compartments”).

218

219 As regards what is to be measured, the EEA definition of State refers to “condition”,  
220 suggesting that relevant information in this category is concerned with status at a given point  
221 in time. Similarly, the PSR/E and PSR frameworks envisage a static measure reflecting the  
222 stock of the quantity or quality of environmental resources. Such measures may be useful in

223 conveying readily accessible messages about trends over time. However, in terms of relating  
224 this category to others in the framework, information on change in State would be needed.  
225 Indeed, the PSIR and FDES frameworks are concerned with such change (and in this sense  
226 resource flows) in the environment, although FDES allows for other perspectives given its  
227 recognition that there may be a direct link between changes in stocks and environmental  
228 impacts (UN, 1984, paras. 46 & 47), while the relevant S-RESS category encompasses both  
229 “effects” on the natural environment and to stock of available resources.

230

## 231 2.5 Impact category

232

233 In DPSIR, impacts are expressed in terms of the effects of “environmental quality” on both  
234 social and ecological systems. Thus, in Cave *et al.* (2003) and Turner *et al.* (1998) impacts  
235 include both ecosystem and human welfare effects, although in the former case they focus on  
236 ecosystem effects such as “reduced water quality”.

237

238 The inclusion of effects on ecosystems obscures the boundary between this category and the  
239 State category. Referring to the eutrophication example, the envisaged distinction in DPSIR  
240 could be between eutrophication as the relevant State and its consequences, such as the effects  
241 of hypoxia on particular species, as the Impact. However, this division would introduce an  
242 artificial distinction since ecosystem changes attributable to eutrophication would also fall in  
243 the State category since they reflect the “condition of different environmental compartments  
244 and systems”. Thus, ecosystem changes attributable to eutrophication could be treated as a  
245 manifestation of Impact or an aspect of State. Alternatively, if State is taken to represent the  
246 availability of eutrophying agents, then Impact would embrace both the extent of

247 eutrophication and its consequences, and thus would be seeking to encapsulate a wide range  
248 of information, particularly given its inclusion of consequences for social systems.

249

250 A further issue arises from including effects on social and ecological systems in the same  
251 category. This conflates distinct concepts, with their own systems of measurement, and  
252 potentially disguises the relationships between them. Moreover, defining the ultimate  
253 consequences of anthropogenic ecosystem change in this way complicates comparison with  
254 the human activities that gave rise to them and can thus hamper Response decisions.

255 Separating the two forms of Impact to isolate the effects on social systems would make  
256 information in this new category comparable with that in the Driver category. With  
257 appropriate economic measurement in these categories there is thus direct correspondence  
258 with the elements required for cost-benefit analysis. To illustrate with the eutrophication  
259 example, say that the use of fertilisers increases agricultural yields and so welfare by A but  
260 leads to ecosystem Impacts, measured in biophysical terms as B, which generate external  
261 welfare costs of C. Only A and C are directly commensurable and respectively represent the  
262 benefits and costs of fertiliser use.

263

264 Support for the notion of isolating impacts on social systems in a separate information  
265 category can be drawn from predecessor frameworks that recognise effects on social systems:  
266 PSR/E and PSIR provide separate categories for such effects. Similarly, in their applications  
267 of DPSIR, both Bowen and Riley (2003) and Atkins et al. (2011) recognise the need for  
268 separate recognition of welfare effects.

269

270 2.6 Response Category

271

272 This category is the most uncontroversial in that the frameworks summarised in Table 2 share  
273 a common conception of information dealing with human reaction to environmental  
274 status/change. Furthermore, none confines the category to management action at a particular  
275 level represented by the other categories. Throughout, categories corresponding to Response  
276 include, for example, policy measures to reduce Drivers as much as to remediate or mitigate  
277 Impacts.

278

## 279 2.7 Omitted Categories

280

281 Having considered each of the categories included in DPSIR, the question arises as to whether  
282 these omit types of information covered in the other frameworks. In this context, the FDES  
283 framework includes “natural events” in its “action” category but this is on the basis that  
284 human activities can contribute to certain such events (UN, 1984, para. 41). The relation  
285 between these activities and events can be catered for through DPSIR’s Driver and State  
286 categories; a separate category is unnecessary. Nevertheless, this concept of “natural events”  
287 highlights the broader issue that relationships within ecosystems may be moderated or  
288 exacerbated by natural conditions or their variability that are exogenous to DPSIR categories,  
289 i.e. not the result of human activity. While S-RESS allows recognition of “natural forces” as  
290 a source of stress on the environment, alone or in conjunction with anthropogenic stresses  
291 (Friend, 1979, p.76), none of the frameworks allocates a category to information on such  
292 conditions/variability. Rather, it is implied that these forces are manifested through changes  
293 in the relationships between categories, i.e. in the application of the accounting framework.  
294 They do not constitute a separate factor in the chain of causal relationships that set the scope  
295 of the information categories.

296

### 297 **3. Modifying DPSIR – the DPSWR Framework**

298

299 The modifications of the DPSIR framework respond to the definitional and conceptual  
300 limitations, and incorporate lessons learned from the examination of precedents, as discussed  
301 above. In overview, the number of categories is retained to ensure they are the minimum  
302 necessary for representing relevant information and their scope is aligned with social or  
303 ecological systems. Table 3 summarises the resulting information category definitions  
304 employed in the DPSWR framework and the relationships between them are depicted in  
305 Figure 1. This figure also highlights their configuration as regards social or ecological  
306 systems, reflecting the distinction between phenomena associated with human agency and  
307 those associated with conditions or processes in ecosystems. Thus, the framework isolates  
308 categories of information relating to social systems.<sup>3</sup>

309

310 Each of the DPSWR categories is discussed below within a spatially-framed conceptual  
311 framework involving serial input-output relationships represented by notional coefficients. It  
312 is emphasised that this is a means of making explanations more concrete and conveying  
313 conceptual relationships in a more formalised setting so as to highlight issues in  
314 measurement; it does not purport to constitute a model. In reality, the relation between two  
315 categories may be non-monotonic (e.g. exhibiting threshold effects) and/or dependent on a  
316 variable from another category. Nevertheless, this form of description may be used as a basis  
317 for identifying elements relevant to specifying models and inform decisions on relevant scales  
318 and the selection of variables appropriate to specific applications of the framework.

---

<sup>3</sup> The form of measurement for each information category is not prescribed by the framework but the system division denotes a broad distinction between types of available measurement. Categories associated with ecological systems involve the objective observation of nature; those associated with social systems may be represented by physical measures or socially constructed measures of 'value'. However, a common criterion in the selection of measures is that they be appropriate to the objective of the analysis for which the framework is employed – in accounting theoretic terms, the framework is a means of processing information, which entails a purpose orientation (Chambers, 1966: 162).

319 Furthermore, the specific application will influence for which categories information is  
320 required; not all categories will necessarily be pertinent for every analysis.

321

### 322 3.1 Driver - Pressure

323

324 The key modification in the Driver category is to focus it on human activities that give rise to  
325 Pressure on natural systems. Thus, the relation between the two is made more direct,  
326 overcoming DPSIR's apparent concentration on large scale and potentially long-term changes  
327 in social systems which are at some remove from specific activities that precipitate ecosystem  
328 change. However, scope is offered for recognising these broader changes as well as the  
329 immediate activities associated with Pressures to facilitate studies concerned with trends in  
330 fundamental factor such as population and consumer choice that influence the level and nature  
331 of activities giving rise to Pressure. This is achieved by focussing the definition of a Driver in  
332 the modified framework on an activity or process within the social system but, where  
333 necessary to accommodate broader scale analysis, allows separation between "immediate  
334 Drivers" (those proximate to Pressures) and "underlying Drivers" (more closely  
335 corresponding to the Driver category in DPSIR). This is similar to the approach of the PSR/E  
336 framework in its definition of "pressures" (Table 2).

337

338 **[Table 3 and Figure 1 about here]**

339

340 In applications involving multiple Drivers, organising information on an economic-sectoral  
341 basis can help direct attention to which parts of the economy are most salient in respect of the  
342 environmental issue at stake. This may be useful in the scoping or prioritisation of analysis in  
343 practice, or in highlighting critical areas for policy development. Similarly, where the

344 analysis is concerned with long-term social trends, organising immediate Drivers in this way  
345 can highlight where such trends are having most effect. A further benefit of sectoral  
346 classification arises where the application is motivated by understanding the range of impacts  
347 that given Drivers have. The sectoral scheme is not prescribed as part of the framework as the  
348 most appropriate level of detail is dependent on the application. In the ELME project, for  
349 example, a high level scheme with nine sectors derived from Eurostat (1999), as shown at the  
350 foot of Table 3, was adequate.

351  
352 In DPSWR, the Pressure category takes on a broad meaning, representing the mechanism or  
353 process that intermediates between human action (Driver) and the relevant ecosystem State.  
354 As such, the category has a more abstract definition than in DPSIR and allows various uses  
355 dependent on the analytical context. This flexibility aids the applicability of the DPSWR  
356 framework to a range of contexts within the constraint of the Driver and State category  
357 definitions being more firmly fixed. Indeed, in certain cases the category may be redundant.  
358 For example, in the case of an emission from some human activity, it is the resulting increase  
359 in concentration of the pollutant in the environment that constitutes the Pressure on the  
360 ecosystem measured in the State category. By contrast, if the human activity directly affects  
361 the ecosystem of interest, e.g. loss of habitat due to dredging activities, there is no need to  
362 specify a Pressure measure. In such cases, the Driver category can still be used to examine  
363 alternative actions, e.g. different methods of dredging and disposal of spoil.

364  
365 To examine the implications for measurement in DPSWR's Driver-Pressure relationship,  
366 consider measures of a range of human activities,  $a_{xi}$ , indexed by the suffix  $x$  ( $x = 1, 2, \dots, X$ )  
367 at location  $i$  ( $i = 1, 2, \dots, I$ ), the set of which is represented by the column vector  $\mathbf{A}_i$  of order  $X$   
368  $\times 1$ . Given the set of activities,  $\mathbf{A}_i$ , the resulting exploitation of environmental resources, e.g.

369 emissions,  $e_{yi}$ , can be connected through notional coefficients,  $\alpha_{yxi}$ , where  $y$  indexes the type  
 370 of exploitation. Thus, for location  $i$ , the set of exploitations may be represented as:

$$371 \quad \alpha_i \mathbf{A}_i = \mathbf{E}_i$$

372 ..... (1)

373

374 where  $\alpha_i$  is a  $Y \times X$  matrix the elements of which are the respective coefficients and  $\mathbf{E}_i$  is a  
 375 column vector of order  $Y \times 1$ , the elements of which are the measures of total environmental  
 376 exploitation of type  $y$  from activities in location  $i$ ,  $e_{yi}$ . For example, if  $y = 1$  represents the  
 377 emission of nitrogen species to water and  $x = 1$  represents the amount of fertiliser used for a  
 378 given period,  $\alpha_{11i}$  represents the rate at which this pollutant is produced by this activity at  
 379 location  $i$ . Furthermore, the first element in  $\mathbf{E}_i$  is the total emission of nitrogen species to  
 380 water from activities at location  $i$  in that period, which would also include, for example,  
 381 atmospheric deposition of nitrogen from activities involving combustion at this location.  
 382 Thus, equation (1) encapsulates the essential features of the Driver-Pressure relationship and  
 383 highlights the significance of the coefficients,  $\alpha_{yxi}$ , which can be used to represent various  
 384 aspects of the relationship. Variation in a given coefficient across locations,  $i$ , may be used to  
 385 represent differences in natural transport processes, i.e. the extent to which a given Driver  
 386 activity results in Pressure over a given period, while variation over time at a given location  
 387 could be used to represent information on decoupling between Driver activities and Pressures,  
 388 for example due to management practices, or a closer coupling due to exogenous changes.

389

390 The scale selected to define location and the measures of activity depend on the application.  
 391 For example, a national government may restrict its attention to activities within its territory  
 392 or jurisdiction while a regional environmental authority, transcending national boundaries, is

393 more likely to be concerned with activities at a geographical scale appropriate to a specific  
 394 ecosystem change and thus, potentially, with activities in various countries.

395

### 396 3.2 Pressure - State

397

398 In common with other frameworks, State is defined in general terms by reference to  
 399 “attributes” (see Table 3) to allow flexibility in application but is most directly comparable to  
 400 the S-RESS and FDES frameworks scope in explicitly providing for both static and dynamic  
 401 measures of environmental conditions. To illustrate by reference to the input-output analysis,  
 402 consider static measures of the relevant conditions at location  $j$ ,  $c_{zj}$ , where  $z$  ( $z = 1, \dots, Z$ )  
 403 indexes the type of environmental condition. As noted above, such static measures may be  
 404 compared at different points in time to identify trends but by themselves they do not generally  
 405 support analysis of the relationships between State and its adjacent categories, which involve  
 406 measurement of change over periods of time. Therefore, where the application involves  
 407 relating information across categories, measures of change in conditions,  $d_{zj}$ , across periods  
 408 comparable with those used for measurement in other categories, are required. Where the  
 409 elements of  $\mathbf{D}_{ij}$  (a  $Z \times 1$  vector) are the measures of change in different types of condition  $d_{zj}$ ,  
 410 and  $\beta_{ij}$  is a  $Z \times Y$  matrix, the elements of which are coefficients linking respective Pressures  
 411 arising from activities at  $i$  to change in environmental conditions at  $j$ , the relationship between  
 412 the Pressure and State categories can be represented by:

$$413 \quad \beta_{ij} \mathbf{E}_i = \mathbf{D}_{ij} \quad \dots\dots\dots (2)$$

415

416 For example, if  $z = 1$  is taken to represent eutrophication then  $d_{1j}$  as the first element in  $\mathbf{D}_{ij}$  is a  
 417 measure of the change in eutrophication at location  $j$  attributable to Pressures from human  
 418 activities at location  $i$ .

419

420 While the coefficients  $\beta_{ij}$  in equation (2) represent purely natural processes, they also  
 421 represent information on human-ecosystem interactions in that they reflect to what extent a  
 422 unit of Pressure from Driver activities at a given location,  $i$ , is translated into State change at  $j$ .  
 423 Comparing coefficients across different locations  $j$  indicates the physical distribution of  
 424 environmental consequences from those activities. This is an anthropocentric orientation and  
 425 applications motivated by State change at a specified location  $\mathbf{D}_j$  resulting from activities at  
 426 multiple locations can be represented by (notional) summation:

$$427 \quad \mathbf{D}_j = \sum_i \mathbf{D}_{ij} \quad \dots\dots\dots (3)$$

429

430 Indeed, in practice, it is the values in  $\mathbf{D}_j$  that are most likely to be observed initially, and  
 431 analysis is required to identify which Driver locations,  $i$ , are most material in terms of relative  
 432 values of  $e_{yi}$  and  $\beta_{ij}$  coefficients across locations.

433

### 434 3.3 State - Impact/Welfare

435

436 Modifications of DPSIR's State and Impact categories are interrelated. The underlying motive  
 437 is to isolate the effects of ecosystem change in terms of social systems from the ecosystem  
 438 changes resulting from Pressures. Thus, the Impact category is redefined in DPSWR to cover  
 439 only information relating to social system effects (see Table 3) and the ecosystem changes  
 440 which DPSIR would have treated as Impacts are dealt with in the State category so that this

441 category encompasses all ecosystem changes other than those which constitute Pressures;  
 442 where the boundary lies between them depends on the application context, as discussed in  
 443 section 4 below. The renaming of the Impact category as Welfare signifies this change in  
 444 scope and improves communication with natural scientists who can find it challenging to  
 445 employ the word “impact” as exclusively relating to social systems and instinctively use it to  
 446 also encompass ecosystem change.<sup>4</sup> The DPSWR nomenclature supports this usage in that  
 447 “impact” can refer to State and/or Welfare changes.

448  
 449 By designating a separate category for information relevant to assessing Welfare effects of  
 450 ecosystem change, the DPSWR framework draws a clear distinction that makes the  
 451 relationship between them more explicit and highlights the need for human agency in linking  
 452 State to Response, since Response is contingent on human perception of values associated  
 453 with State. Moreover, this distinction supports comparison of the human causes and effects of  
 454 environmental change: the Driver category identifies activities the economic benefit of which  
 455 can be assessed next to the costs in the Welfare category. Furthermore, the distribution of  
 456 benefits and costs can be revealed in the measurement scheme. Continuing the input-output  
 457 representation above, the change environmental conditions at location  $j$  can be notionally  
 458 translated into welfare effects at location  $k$ ,  $w_k$ , through the  $K \times Z$  coefficient matrix,  $\gamma_{jk}$  to  
 459 yield a vector of welfare effects,  $\mathbf{W}_k$ :

$$\gamma_{jk} \mathbf{D}_j = \mathbf{W}_k \quad \dots\dots\dots (4)$$

462 The locations,  $k$ , thus represent where the costs of environmental degradation are borne, while  
 463 the locations,  $i$  (equation 1) represent those where the activities giving rise to that change took  
 464 place and thus where benefits arise. Any mismatch between the two reflects different

---

<sup>4</sup>I am grateful to Laurence Mee for this insight.

465 distributions of benefits and costs and where these cross institutional, e.g. national, boundaries  
466 this raises the question of equity.

467

468 Welfare effects encompass a range of human values as regards environmental change,  
469 including non-use value motivated by bequest, existence or ‘anthropocentric intrinsic’ values,  
470 that are potentially capable of representation with a common (monetary) metric (Turner,  
471 1999). Such monetary representation enables lost benefits, or costs, to be aggregated across  
472 types of environmental change and their location, and to be compared with monetary  
473 representation of the benefits associated with the Driver activities from which the Welfare  
474 effects derive, i.e. cost-benefit analysis. Similarly, such analysis might be applied in  
475 assessing the cost of Response actions against the benefit of environmental improvement.  
476 However, in common with the other DPSWR categories, the Welfare category is defined  
477 conceptually rather than in terms of its measurement base. Consequently, in applications  
478 where there is a lack of reliable valuation data, indicator-based measures of relevant criteria,  
479 e.g. an increase in morbidity or reduction in employment attributable to State change, may be  
480 used to represent the Welfare category. With appropriate weightings, such indicators can be  
481 combined to represent multiple criteria in a single measure (i.e. multi-criteria analysis, OECD,  
482 1989: 19) that can substitute for an aggregate monetary value for those criteria.

483

484 Through its use of the welfare concept, the DPSWR framework gives a primary role to  
485 assessing changes in ecosystem services in terms of the effects on humans, thus aligning it  
486 with the information needs of various decision making contexts (Fisher et al., 2009) and  
487 policies such as the Marine Strategy Framework Directive (EU, 2008:29). However, the  
488 content of the Welfare category can be adapted to other decision-frames such that ‘welfare’ is  
489 given a broader meaning as ‘what matters’ through human instrumentality, i.e. human agency

490 in defining decision criteria. For example, where cost-benefit analysis is rejected on  
491 economic-theoretic grounds (Gowdy, 2004), under conditions of uncertainty (Perrings, 1991),  
492 or, more generally, as an inappropriate basis for environmental decision-making (O'Neill,  
493 1997; Sagoff, 2004), the category could be used to encapsulate information on minimum  
494 acceptable levels of ecosystem provision in accordance with the precautionary principle or an  
495 ethically motivated desire to maintain such a level 'for its own sake' (i.e. 'anthropocentric  
496 intrinsic value' in the classification scheme of Turner, 1999:35).

497

#### 498 **4. Illustration in Marine Ecosystems**

499

500 To test the feasibility of the DPSWR framework, natural science teams working in the ELME  
501 project were requested to apply the framework in identifying key Drivers and Pressures for  
502 the environmental issues with which they were concerned, using the definitions and a slightly  
503 extended version of the notes shown in Table 3. Examples derived from their responses are  
504 presented in Tables 4 and 5 to illustrate the application of the framework in the marine  
505 context. These examples are selected as they involve differing treatments of the same class of  
506 item: seagrass loss in the Mediterranean (Table 4) includes various forms of pollution as  
507 Pressures while that of chemical pollution in the Northeast Atlantic (Table 5) takes pollution,  
508 rather than its consequences for ecosystems, as the State of interest. This reflects the scope of  
509 work of the respective teams and illustrates that the treatment of an item as a Pressure or a  
510 State variable depends on the application context. A similar flexibility is apparent in policy.  
511 The EU's Marine Strategy Framework Directive (EU, 2008) prescribes descriptors of "good  
512 environmental status" (in Annex I). Some of these directly correspond to State (e.g. no.1  
513 "biological diversity is maintained"), while others may be seen more generally as Pressures

514 (e.g. no.11 “introduction of energy ... is at levels that do not adversely affect the marine  
515 environment”).

516

517 **[Tables 4 and 5 about here]**

518

519 Despite the difference in the role of pollution in these cases, both are able to align biophysical  
520 phenomena relevant to their scope with Driver and Pressure categories. However, this involves  
521 different approaches to the classification of Pressures. In the seagrass case there is a range of  
522 types of Pressure (physical, chemical and biological), while for the chemical pollution case  
523 this category is used to organise Drivers according to the route of transmission of the  
524 pollutants to the marine environment, reflecting the significance of different transport  
525 processes as contributors to State change.

526

527 The identification of Drivers and their sectoral classification by the respondents were  
528 consistent with the concepts used in the design of the framework despite the relatively simple  
529 definition of Driver. Indeed, there is evidence of a flexible interpretation in the responses in  
530 that static physical structures such as shipping infrastructure were seen as Drivers although  
531 the “activity or process” with which they are associated is their presence, i.e. “being there”.

532

## 533 **5. Concluding Remarks**

534

535 The DPSWR framework involves a number of modifications to DPSIR which seek to  
536 improve the clarity of information category definitions and establish a conceptual foundation  
537 for each category that supports its linkage with other categories. By defining Drivers in terms  
538 of human activities, a direct link to Pressures is enabled; by expanding the State category to

539 encompass changes in conditions establishes a link between Pressures and changes in human  
540 Welfare over a period of time; and by separating such Welfare changes from State changes  
541 the boundary between social and ecological systems and the interaction between them are  
542 clearly marked. As a result, the DPSWR information categories relating to social systems  
543 highlight the link between the human activities that give rise to environmental change,  
544 whether as a result of actions in the Driver or Response categories, and the effect of such  
545 change on humans. Thus, the framework isolates information relevant to the requirements of  
546 cost-benefit analysis and other decision frames insofar as these reflect human, rather than  
547 intrinsic, values.

548

549 Modifying DPSIR in the ways embodied in the DPSWR framework was found to be feasible  
550 in the project that motivated this study and imports advantages to applications in terms of  
551 supporting the identification of representative variables for each category and comparability  
552 across studies. Furthermore, the conceptual input-output analysis employed here to represent  
553 relationships between categories indicates types of information needed to fully account for  
554 interactions between human and ecological systems.

555

556 In common with previous socio-ecological accounting frameworks, DPSWR is less  
557 prescriptive than institutional accounting frameworks in its definition of information to be  
558 provided but is more comprehensive in its scope. This allows flexibility in its application but  
559 results in a richer, more integrated portrayal of human-ecosystem interactions, as is apparent  
560 in reconciling this framework with those employed in institutional accounting (Table 1). In  
561 the case of macroeconomic accounting, the primary link with DPSWR is through limiting the  
562 State category to a scale in accordance with national boundaries, for example UNSD (2012)  
563 defines the scope of “environmental assets” by reference to “the economic territory over

564 which a country has control ... including waters and sea-beds within a country's Exclusive  
565 Economic Zone" (p.124). Thus, scope is restricted in accordance with anthropocentric criteria  
566 which may not correspond to the scales required to fully capture information on ecological  
567 change as envisaged in DPSWR. In corporate level accounting, the current practice (as  
568 illustrated by the examples in Table 1) is to limit accountability in reporting and management  
569 information systems to the immediate results of the entity's Driver activities, its emissions or  
570 resource use, corresponding to Pressures in DPSWR. However, where the entity operates at  
571 multiple locations, information on Pressures aggregated at the corporate level may be of  
572 limited usefulness.

573

574 While socio-ecological accounting frameworks such as DPSWR offer the potential for more  
575 detailed policy-relevant information than institutional frameworks, they have thus far been  
576 concerned primarily with the definition of relevant types of information. Their further  
577 development requires consideration of how to incorporate information on the temporal lags  
578 between measures of different categories (e.g. *when* Drivers are manifested in Welfare  
579 changes) and the degree of uncertainty in the relationship between information categories, as  
580 well as how these relationships may be affected by other changes in future (the importance of  
581 which is noted by Heal, 2007, in the context of national accounting). These factors could be  
582 material to decisions based on applications of the DPSWR framework.

583

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585

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592

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704

705 Table 1: A typology of environmental accounting frameworks

Perspective	Economic scope/ Accounting entity	Environmental scope	Measurement	Examples
Institutional	Macroeconomic/ Geo-political entity (typically a country)	Attributable (national) resource stocks and flows	Monetary, e.g. environmentally adjusted national income, “environmental asset” values <sup>a</sup>	Satellite national accounts (UNSD, 2003: 450, 2012) Sustainability indicators (Hamilton and Clemens, 1999; Pearce and Atkinson, 1993)
			Physical units, e.g. physical flows between economy and environment	Satellite national accounts (UNSD, 2012)
			Total resource use	Geographical area required to meet entity’s needs
	Microeconomic/ Corporate entities (e.g. firms)	Attributable natural resource use (as source or sink)	Physical units <sup>c</sup>	Sustainability reporting guidelines (GRI, 2011) Internal ecological accounting (Schaltegger and Burritt, 2000: Ch.11)
Socio- ecological	“Mesoeconomic”/ Scope required to account for social system interactions relevant to the environmental scope	As specified for the purpose of the analysis	Indicators representing status or change by information category	DPSIR (EEA, 2007)

706 <sup>a</sup> For an overview, see Lawn, 2007.707 <sup>b</sup> Ecological footprinting has been applied to sub-national entities, e.g. Greater London (BFF, 2002).708 <sup>c</sup> The use of monetary measures in corporate environmental information systems and reporting (e.g. applying  
709 external damage estimates to emissions or costing replacement/amelioration) is conceivable but has not been  
710 widely taken up in practice.

711 Table 2: Summary of socio-ecological accounting framework information categories and their content  
712

Framework	Information Categories				
	Driver	Pressure	State	Impact	Response
DPSIR (EEA, 2012)* (EEA, 1999)**	Driving force** <i>social, demographic and economic developments in societies and the corresponding changes in life styles, overall levels of consumption and production patterns</i>	Pressure** <i>developments in (the) release of substances (emissions), physical and biological agents, the use of resources and the use of land</i>	State (indicator)* <i>indicator of condition of different environmental compartments and systems in physical, chemical or biological variables</i>	(Environmental) impact* <i>impacts on human beings, ecosystems and man-made capital resulting from changes in environmental quality</i>	Response** <i>responses by groups (and individuals) in society, as well as government attempts to prevent, compensate, ameliorate or adapt to changes in the state of the environment</i>
PSIR (Turner 2000; Turner et al., 1998)	Socio-economic drivers <i>urbanisation and transport/trade, agricultural intensification/land-use change, tourism and recreation demand etc.</i>	Environmental pressures <i>land conversions and reclamation, dredging, aggregates and oil and gas extraction, waste disposal etc.</i>	Environmental “state” changes <i>changes in ... fluxes across coastal zones, loss of habitats and biodiversity etc.</i>	Impacts <i>consequential impacts on human welfare via productivity, health, amenity and existence value changes</i>	Policy response options -
PSR/E (Schulze and Colby, 1994)	Pressures		Direct pressures <i>biophysical inputs and outputs that may exert immediate stress on ecosystems</i>	State <i>Ambient conditions and trends Valued environmental attributes</i>	Societal response <i>purposeful actions to address ... ecological, human health or welfare changes or impacts that are considered undesirable</i>
	Underlying <i>social and technological forces that ... drive economic activity</i>	Indirect <sup>a</sup> <i>human activities related to ... improvement of human welfare</i>			
Effects <i>Relationships between two or more variables within any of the pressure, state and response categories</i>					
PSR (OECD, 2003; 1993)	Pressure		State		Response
	Indirect <i>human activities which lead to proximate pressures</i>	Proximate <i>pressures directly exerted on the environment e.g. emissions</i>	<i>quality of the environment and the quality and quantity of natural resources</i>		<i>actions to mitigate, adapt to or prevent human-induced negative impacts on the environment</i>
FDES (UN, 1991; 1984)	Action <i>social and economic activities, natural events</i>		Impact <i>environmental impacts of activities/events</i>		Reaction <i>responses to environmental impacts</i>
S-RESS	Stressor	Stress	Environmental response		Collective and individual

(Friend, 1979)	<i>activities with the potential to degrade the quality of the natural environment, to effect (sic) the health of man, to threaten the survival of species, to place pressure on non-renewable resources, and to deteriorate the quality of human settlement</i>	<i>elements that place pressures on, and contribute to the breakdown of, the natural and man-made environment</i>	<i>observed effects of stress upon natural and man-made environments</i> Stock <i>available resource (cf. flows in other measures)</i>		responses <i>man's reaction to environmental changes</i>
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713  
714  
715

<sup>a</sup> Also includes “natural processes and factors” which may act alone or together with human actions to create biophysical pressures.

716 Table 3: Definition of DPSWR information categories  
717

Information Category	Definition	Commentary
Driver	An activity or process intended to enhance human welfare.	<ul style="list-style-type: none"> <li>▪ Organising activities into economic sectors assists in directing attention to the most salient areas of the economy.<sup>a</sup></li> <li>▪ Where necessary the category can be split between: <ul style="list-style-type: none"> <li>- Immediate Drivers: activities proximal to at least one Pressure.</li> <li>- Underlying Driver: population, economic, social and technological factors that influence the level/nature of Immediate Drivers.</li> </ul> </li> </ul>
Pressure	A means by which at least one Driver causes or contributes to a change in State.	Thus a pressure is a link between a Driver and a change in environmental State, effectively therefore the agent of change. Generally, it is a by-product or an unintended consequence of the Driver activity/process.
State (change)	An attribute or set of attributes of the natural environment that reflect its integrity as regards a specified issue (or change therein).	<p>This definition allows flexibility so that the information or measure used can be tailored to the precise circumstances that are relevant. However, often the most useful information will:</p> <ul style="list-style-type: none"> <li>▪ relate to the extent to which a system has been subject to disturbance, particularly in terms of ecosystem functionality, and</li> <li>▪ reflect changes in State over time.</li> </ul> <p>Natural (i.e. non-anthropogenic) variability may influence the effect of Pressures on State or change in State.</p>
Welfare	A change in human welfare attributable to a change in State.	<ul style="list-style-type: none"> <li>▪ “Change” allows for enhancement but generally we are concerned with diminution in welfare.</li> <li>▪ Welfare is not only affected by changes in use values; it can be affected by changes in nonuse values that people hold (e.g. in respect of general ecosystem functionality or the viability of particular species).</li> </ul>
Response	An initiative intended to reduce at least one Impact (State or Welfare change).	In this sense “initiative” is an action that would not have been taken in the absence of an effect on Welfare. It may operate through influencing any of the above but with the intention to ultimately reduce such an effect.

718 <sup>a</sup> For example, the following sectors were used in the ELME project (based on Eurostat, 1999): Agriculture,  
719 Energy, Fisheries & Aquaculture (including extraction of biological resources), Household (individual  
720 consumption), Industry, Tourism & Recreation, Transport, Urbanisation & Coastal Development.  
721

722 Table 4: Summary of drivers and pressures resulting in loss or degradation of seagrass beds in the Mediterranean  
 723

<b>Pressure</b>	<b>Driver</b>	<b>Sector<sup>a</sup></b>
Nutrient (nitrogen & phosphorus) release	Fertiliser use; Intensive livestock management	Agriculture
	Release of sewage after low level treatment; Use of inadequate sewage systems (resulting in leakage and storm water overflows)	Urbanisation
	Aquaculture (discharge of waste food, faecal and dead animal waste)	Fisheries
Mechanical disturbance	Dredging and spoil disposal; Propeller and anchor damage from shipping	Transport
	Laying of submarine pipelines and cables; Land claim	Urbanisation
	Use of mobile gears (e.g. trawling and dredging)	Fisheries
	Aggregate extraction	Resource Extraction
	Anchoring of pleasure craft	Tourism & Recreation
Introduced organisms	Accidental release of organisms from aquaria (e.g. <i>Caulerpa taxifolia</i> )	Tourism & Recreation
	Use of mobile fishing gears (leading to spread of <i>Caulerpa</i> )	Fisheries
Contaminants	Land claim using contaminated landfill; Sewage sources, as above	Urbanisation
	Industrial processes (discharge of untreated wastes)	Industry
	Dredging/spoil disposal; Shipping waste disposal (e.g. oil and fuel)	Transport
Physical oceanographic change (increase in turbidity)	Aggregate extraction	Resource Extraction
	Beach replenishment using terrigenous material	Tourism & Recreation
	Land claim; Coastal defence construction/modification; Discharge of sewage particulates	Urbanisation
	Land management, e.g. deforestation (increased deposition of sediment)	Agriculture
	Aquaculture (discharge of waste food, faecal and dead animal waste)	Fisheries
Direct removal	Harvesting for use in agricultural fertiliser production	Fisheries
	Removal to “improve” aesthetics of tourist beaches.	Tourism & Recreation
Anthropogenic structures (producing shading)	Shipping infrastructure: ports, harbours, jetties	Transport
	Land claim; Protection of urban areas/infrastructure through coastal defences	Urbanisation
Chemical oceanographic change	Water abstraction	Extraction
	Untreated waste discharge (increasing chemical oxygen demand)	Industry
	Untreated sewage discharge (increasing chemical and biological oxygen demand)	Urbanisation
Deposition of physical material	Laying of submarine pipelines and cables; Sewage sludge disposal	Urbanisation
	Construction/maintenance of artificial reefs	Tourism & Recreation

724 <sup>a</sup> “Fisheries” includes biological resource utilisation and aquaculture; “Transport” includes maritime traffic and construction/preservation of navigable routes

725 Table 5: Summary of drivers and pressures resulting in chemical pollution in the Northeast Atlantic

726

<b>Pressure</b>	<b>Driver</b>	<b>Sector</b>
Chemical products discharged into rivers, coastal waters or offshore	Disposal of urban waste; Release of sewage after low level treatment; Use of inadequate sewage systems (resulting in leakage and storm water overflows); Coverage of land with urban space and roads (producing run-off water)	Urbanisation
	Use of domestic chemicals and pharmaceuticals (including hormones) present in wastewater	Household
	Industrial processes	Industry
	Aquaculture (leading to release of chemotherapeutics, hormones)	Fisheries
	Shipping waste disposal; Release of antifouling agents	Transport
Atmospheric emissions deposited in rivers, coastal waters or offshore	Use of pesticides/herbicides (contaminating run-off water)	Agriculture
	Industrial processes	Industry
	Operation of terrestrial vehicle and ship engines	Transport
	Combustion for electricity generation	Energy
Chemicals accidentally released into rivers, coastal waters or offshore	Oil and gas refinery processes; Flaring by oil and gas production platforms	Energy
	Use of pesticides/herbicides	Agriculture
	Industrial processes	Industry
	Transport of oil/chemicals	Transport
	Oil production processes	Energy

727

728

Figure 1: Summary of DPSWR definitions and relationships

